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1 **Variation in suspended sediment yield across the UK – a failure of the concept and**
2 **interpretation of the sediment delivery ratio**

3

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10

11 **Abstract**

12 The sediment delivery ratio (SDR) has been a common approach developed to understand
13 change in sediment yield and flux through a catchment. In this study we propose that the
14 underlying concept of the sediment delivery ratio is flawed for a number of reasons: its linear
15 extrapolation is physically meaningless; there is no evidence of the magnitude of storage
16 required by the SDR approach on annual to decadal timescales; and the SDR approach
17 assumes suspended sediment transport is conservative yet it is known to undergo both loss
18 and production in-channel.

19 This study considers the sediment yield from 192 UK catchments from 1974 to 2010
20 for catchment areas between 4 and 9948 km² and shows that linear extrapolation of the SDR
21 approach overpredicts source terms and underpredicts fluxes for large catchments. The SDR
22 approach hides a range of behaviours of suspended sediment flux within catchments with
23 patterns of net deposition, net increase or no change all apparent in UK catchments. The

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24 approach proved to be self-correlated which meant that it can result in spurious correlations
25 when compared to catchment area. The change in yield with catchment area can be just as
26 well understood as a change in sediment supply from channels rather than as a change in
27 delivery from hillslope sources. We propose that suspended sediment flux change with
28 catchment area be modelled as a more physically-meaningful Gompertz function (step
29 function) rather than using the traditional SDR approach.

30

31 **Keywords:** suspended sediment flux; SDR; particulate organic matter; rivers

32

33 **1. Introduction**

34 The sediment delivery ratio (SDR) is a common approach used to explain changes in
35 sediment flux through a catchment (Roehl, 1962; Burt and Allison, 2009). The concept was
36 developed because of the need to link observed erosion at field or hillslope scale to sediment
37 yields in larger catchments and to then predict sediment yield and flux at ungauged sites (e.g.
38 Bouraoui and Dillaha, 1996). Here, flux is defined as the mass of suspended sediment per
39 year (tonnes/yr) and the yield is the flux per unit area per year (tonnes/km²/yr; yield is often
40 referred to as export). White (2005) described the sediment delivery ratio as a
41 conceptualisation of the erosion-transport-deposition cycle. The black-box nature of the
42 interpretation of SDR has led to criticism (e.g. Trimble and Cosson, 2000) and its subsequent
43 replacement by sophisticated deterministic models (e.g. Borah and Bera, 2004) which
44 nonetheless remain opaque due to the complexity of their linkages and difficulties of
45 parameterisation.

46

47 *1.1 Questioning the geomorphological basis for a Sediment Delivery Ratio*

48 Here we are interested in the SDR approach as it pertains to suspended sediment flux. We
49 argue that, for a number of reasons, the use of the SDR approach could have misled the
50 interpretation of suspended sediment flux. Firstly, the term SDR implies that the major reason
51 for sediment yield to decrease with increasing catchment area is that delivery of sediment
52 gets less and less, i.e. sediment is not delivered to lower reaches of the basin but rather is lost
53 to storage (Fryirs, 2013). This can be seen in the review of SDR given by Walling (1983)
54 where the spatial lumping of the SDR concept has been recognised but has been interpreted
55 as difference in delivery across the catchment, with each unit of the catchment having its own
56 SDR. Several authors have then been forced to explain a decline in sediment yield with
57 increasing catchment area as due to a decline in connectivity as catchment area increases
58 (Delmas et al., 2012; Fryirs, 2013) and, although they talk of changing erosion rates through
59 a catchment, this is not accounted for. Relying on loss to storage to explain changes in
60 sediment yield with catchment area implies large sinks, but such large sink sizes are difficult
61 to explain, identify or justify given the known, measured rates of deposition.

62 Worrall et al. (2013) calculated the suspended sediment flux for the UK, for which the
63 SDR would be 10%, although if an empirical rule were applied (Roehl, 1962), then the SDR
64 would be approximately 4%. Based on the average suspended sediment flux from the UK
65 from 1974 to 2010, at an SDR of 10%, then 84 Mtonnes/yr would have to be stored
66 somewhere in the UK each year, this is equivalent to an average 344 tonnes/km²/yr stored for
67 each km² of the UK. This equates to an annual net deposition across the UK of 0.23
68 mm/ha/yr based upon a bulk density of 1.5 tonnes/m³.

69 Sediment budgets are often estimated for small headwater catchments; for example,
70 Walling et al. (2002) studied lowland agricultural catchments in the UK of between 0.96 and
71 2.67 km² and found fluxes at the catchment outlets were between 40 and 209 tonnes/km²/yr
72 with SDR between 14 and 27%. The headwater catchments used by Walling et al. (2002)

73 were described as typical of lowland Britain yet the in-field storage, and storage between
74 field and channel (an SDR of between 14 and 27% equates to between 72 and 86% sediment
75 storage), were still not sufficient to achieve the amount of deposition required at the national
76 scale.

77 If sediment storage in “typical” headwater catchments is not sufficient, are there additional
78 processes operating at larger scales that could remove an even greater percentage of
79 suspended sediment flux? Overbank sedimentation occurs in the lower courses of typical UK
80 rivers, but estimates of overbank sedimentation give very low rates compared to those
81 required. Walling et al. (1999) estimated overbank sedimentation for the Yorkshire Ouse
82 (3313 km² - flux at outlet: 49 ktonnes/yr) as 30% of the outlet flux (23% of influent
83 suspended sediment flux) and as 40% of the outlet flux (29% of influent flux) for the River
84 Tweed (4390 km² – flux at outlet: 66 ktonnes/yr). At its maximum, the rate of overbank
85 sedimentation was 6 tonnes/km²/yr. Therefore, compared to sediment storage and SDR
86 required for the UK, the available studies do not show sufficient loss to overbank storage.

87 Overbank sedimentation is not the only storage process operating at large scales in river
88 catchments and there could be in-channel storage. Collins and Walling (2007) give values of
89 in-channel storage as between 18% and 57% of the outlet flux of two UK lowland streams
90 but they note that most of this storage was transient. Indeed, Walling et al. (2002) noted that
91 permanent in-channel storage was only between 2 and 5% of the catchment outlet flux.
92 Therefore, there is no evidence for sufficient in-catchment storage if the sediment delivery
93 problem is only interpreted in such terms as delivery, conveyance and connectivity – there
94 must have been a change in supply. This can be observed in the study of Delmas et al. (2012)
95 where there is a recognition that erosion rate changes with scale from hillslope scale up to a
96 catchment area of 10,000 km² but there was no recognition that the rate of sediment delivery
97 would change with scale, and therefore did not consider there would be any change upon the

98 estimated value of SDR with varying catchment size. In other words a headwater source area
99 will have very different characteristics to one in the lower reaches of a 10,000km² catchment.

100 The fact that slope angles decline towards the lower reaches of most catchments is well
101 known (i.e. the hypsometric curve – Cohen et al., 2008). However, the SDR approach implies
102 that each unit area of the catchment supplies material at the same rate no matter where in the
103 catchment it is and that some areas are disconnected and fail to deliver. Even theoretical
104 approaches to understanding SDR have chosen to take a storage rather than supply approach
105 (Lu et al., 2005). Therefore, we argue that the use of the SDR term has emphasised change in
106 storage rather than change in supply.

107 In general, the SDR is plotted in log-log space which disguises some important but
108 implicit assumptions. When not plotted on the log-scale it becomes apparent that sediment
109 yield actually decreases fastest in the upper parts of a catchment where flood plains are
110 typically small and slopes highest; correspondingly, sediment yield hardly changes in the
111 lower parts of the catchment where flood plains are large and slopes low. The classical
112 picture of the river system is one of erosion in the headwaters, transport in the middle course
113 and deposition in its lower course (Schumm, 1977), but that is not consistent with a sediment
114 delivery that declines fastest in the headwater tributaries of a catchment but varies little
115 further down the channel network. One explanation of this distribution of sediment yield is a
116 common confusion of interpreting sediment yield as sediment flux. Parsons et al. (2006) have
117 indeed argued that sediment delivery is a fallacy when considered via the medium of
118 sediment yield (flux/area) rather than as flux itself.

119

120 *1.2 Conservative and non-conservative sediment*

121 The concept of SDR can be unhelpful if it is interpreted as if suspended sediment were
122 chemically and biologically conservative. The average organic matter content of UK

123 suspended sediment at Harmonised Monitoring Sites (HMS) since 1974 is 33.5% with a 5th
124 percentile of 6% and a 95th percentile of 75% (Worrall et al., 2014). The importance of the
125 non-mineral content of suspended sediment fluxes has been underestimated because it is
126 often quoted as particulate organic carbon content but it is forgotten that naturally occurring
127 organic matter is typically only 45 to 50% carbon and so, if organic carbon concentration is
128 compared to dry mass weight of suspended sediment, its importance is approximately halved.

129 This organic matter content means several problems for how a SDR has been interpreted.
130 Firstly, approximately one third of the suspended sediment may change by independent
131 physical processes such as turnover to dissolved or gaseous phases. For example, Walling
132 and Collins (2008) give eight sediment budgets for agricultural catchments across UK and
133 USA but in none of these sediment budgets is turnover considered or estimated. Secondly,
134 although it is possible under some circumstances for mineral matter to be precipitated in a
135 stream, such effects are rare, whilst the river itself can be a source of particulate organic
136 matter from in-stream flora and fauna independent of erosion processes. Ni et al. (2012) use
137 the SDR approach to predict sediment deposition of particulate carbon in small watersheds
138 without reference to supply varying with scale or the non-conservative nature of the carbon.
139 Similarly, Gaiser et al. (2008) when modelling the effect of soil erosion on carbon
140 sequestration assumed that carbon in eroded sediment no longer influenced CO₂ emissions
141 but rather particulate carbon was either deposited in river channels or lost to the sea.
142 Furthermore, there are sources of suspended sediment that have nothing to do with soils, river
143 banks or erosion, for example, sewage outfalls in the UK typically discharge water with a
144 suspended solid concentrations of between 20 and 30 mg/l which will predominantly be
145 particulate organic matter. Moody et al. (2013) found between 38 and 87% in-stream removal
146 of peat-derived POC over a 10 day period, but even at this rate of loss, or turnover of the

147 organic matter content of suspended sediment would still not be enough to fulfil the storage
148 requirement identified above.

149

150 1.3 *The sediment delivery ratio as a mathematical function*

151 The SDR approach may be more of an empirical than a physical law, however, the
152 “classical” variation in sediment delivery ratio is implicitly assumed to be a power law
153 function of area but this causes some consequences that are rarely interpreted. Firstly,
154 extrapolation to small values of area should tend towards erosion values at the field scale but
155 often do not (e.g. Vanmaercke et al., 2011) which implies that a single power law relationship
156 may not hold true at the lowest values of catchment area even though this was one of the
157 reasons that the SDR approach was developed, i.e. linking universal soil loss equations to
158 catchment management (Wischmeier and Smith, 1965, USDA SCS, 1972, Lim et al., 2005).
159 Equally, the correlation between sediment delivery ratio and catchment area is generally
160 negative and extrapolation of the trend to very large catchments implies that catchments exist
161 with no sediment yield and so, as with the relationship at low values of catchment area, the
162 relationships may not hold true at large values of catchment area. Secondly, the gradients of
163 the relationship between sediment delivery ratio and catchment area vary to the extent that, in
164 some cases, the relationship predicts a decline in actual suspended sediment flux through a
165 catchment. This has been neglected partly because many studies have confused yield and flux
166 (eg. Fryyirs, 2012). We therefore suggest the SDR approach has worked poorly in linking
167 field or hillslope observations to both larger catchment scales and to the discharge of the
168 largest catchments.

169 Finally, a plot of sediment delivery ratio (sediment yield vs. catchment area) will be
170 subject to self-correlation (Kenney, 1982) and therefore presents a spurious relationship that
171 cannot - and should not - be interpreted as meaningful let alone used for prediction. Self-

172 correlation occurs whenever there is a variable common to both the predictor and response
173 variable, in this case the SDR is derived from the sediment yield (the ratio of suspended
174 sediment flux to catchment area) and then compared to catchment area. A variable common
175 to both response and predictor is likely to result in a linear correlation, which is even more
176 likely if the common variable is the dominant source of variance in the response variable. In
177 the case of SDR in the UK, catchment areas vary up to 10000 km² while sediment yield
178 varies up to 1000 tonnes/km²/yr. Self-correlation in the SDR approach has not been tested,
179 yet correlations based upon sediment yield variation with catchment area are commonly used,
180 interpreted and discussed (e.g. Tetzlaff et al., 2013). It should be noted that self-correlation in
181 this context predicts a negative relationship between SDR and catchment area and so positive
182 relationships might be thought to be free of such issues. Such relationships between SDR and
183 catchment area have been observed in a number of studies (e.g. Church and Slaymaker, 1989)
184 and have been associated with, for example, areas of high channel incision. Here, we propose
185 that SDR is self-correlated and should therefore be replaced by measures not prone to
186 spurious correlation.

187 The purpose of this study is to consider the change in sediment yield across the UK as a
188 means of testing the sediment delivery ratio approach and whether the issues raised above can
189 be solved.

190

191 **2 Methodology**

192 This study used the suspended sediment flux data as calculated, corrected and analysed by
193 Worrall et al. (2013), which was based on data from the Harmonised Monitoring Scheme
194 (HMS - Bellamy and Wilkinson, 2001). There are 56 HMS sites in Scotland and 214 sites in
195 England and Wales (Figure 1). Rivers for monitoring were selected as the tidal limit of rivers
196 with an average annual discharge over 2 m³/s; in addition, any tributaries that have an

197 average annual discharge above $2 \text{ m}^3/\text{s}$ are also sampled. These criteria lead to a good spatial
198 coverage of the coast of England and Wales, but in Scotland many of the west coast rivers are
199 too small to warrant inclusion in the HMS. No data were available from Northern Ireland.
200 This study only considered sites where sediment flux monitoring was coincident with flow
201 monitoring, otherwise a flux calculation would be impossible.

202 Among the monitoring agencies, sampling frequencies vary, ranging from sub-weekly to
203 monthly or even less frequently. Annual data were rejected at any site where there were fewer
204 than 12 samples (sampling frequency (f) = 1 month) in that year with the samples in separate
205 months. Flux was estimated using the method of Littlewood and Marsh (2005) who proposed
206 an interpolation technique that accounts for differing sampling frequencies. This approach
207 assumes that each sample taken at a site is equally likely to be representative of an equal
208 proportion of the year as any other sample. The data from the HMS dataset have an
209 inconsistent, and often only monthly, sampling frequency. So as to correct the flux data for
210 sampling bias and to correct the flux estimates to the highest frequency flux estimates
211 analysis of variance (ANOVA) was used to derive all the correction factors after the method
212 and Worrall et al. (2013) and bring all site-year combinations to be equivalent to those with
213 the highest sampling frequency. With respect to the suspended sediment concentration, there
214 were 103162 measurements with 91604 measurements for which there were also measured
215 flow from 270 catchments for all years from 1974 to 2010. An annual suspended sediment
216 flux could be calculated for at least one year in all 270 catchments in the HMS scheme but
217 across the 37 years of the HMS records considered and across the 270 catchment (9472
218 possible catchment-year combinations) a flux calculation was only possible for 6026
219 catchment-year combinations (66%).

220 The estimated suspended sediment fluxes were compared to a range of catchment
221 properties. All suspended sediment fluxes were expressed as a yield (or an export) based

222 upon catchment areas derived for each monitoring point from the CEH Wallingford digital
223 terrain model which has a 50 m grid interval and a 0.1 m altitude interval. The soils of each
224 catchment were classified into mineral, organo-mineral and organic soils based upon the
225 classification system of Hodgson (1997) and the land use was classified into: arable, grass
226 and urban based upon the June Agricultural Census for 2004 (Defra, 2005a). In addition, the
227 number of cattle and sheep in each 1 km² were counted within this census and converted to
228 the “equivalent sheep per hectare” based upon a ratio of 3.1 sheep per cow. A range of
229 hydrological characteristics for each catchment were calculated, these were: the BFI
230 (Baseflow Index - Gustard et al., 1992), the average actual evaporation, the standard average
231 annual rainfall; and the maximum altitude within the catchment. The comparison between
232 suspended sediment fluxes and catchment characteristics was analysed firstly by multiple
233 regression and secondly by logistic regression. For multiple regression, variables were
234 assessed for normality (Anderson and Darling, 1952) and if necessary log-transformed – it
235 did prove necessary for further transformation. The best-fit multiple regression was fitted
236 using both step-up and step-down procedures with variables only retained if they were
237 significantly different from zero at a 95% probability. Quality of fit was assessed using the
238 coefficient of determination (R^2). Binary logistic regression to understand the differences
239 between groups of data found in the analysis. The best-fit binary logistic regression was fitted
240 using maximum likelihood and again variables were only retained if they were significantly
241 different from zero at the 95% probability. The quality of fit was assessed by considering
242 concordance, i.e. the correct classification of the data. For the binary logistic regression the
243 odds ratio was used to assess importance of individual variables. For both multiple regression
244 and binary logistic regression the best-fit line was also chosen based on physical
245 interpretability of the composition of the identified equations.

246 The data can be tested for self-correlation using a method adapted from Vickers et al.
247 (2009). The principle is that in a self-correlated dataset, correlation will be observed even
248 when the values are calculated from values drawn at random. To apply the method, a value of
249 catchment suspended sediment yield is calculated from the catchment's average suspended
250 sediment flux and a catchment area drawn at random from the values within the whole
251 dataset. The new value of sediment yield is paired with the value of catchment area drawn
252 that was used to calculate the sediment yield. This process is repeated a sufficient number of
253 times to create the same number of data pairs as in the original dataset. This random
254 generation of a dataset of equivalent size to the original can be repeated to generate as many
255 random datasets as required - in this case 100 random datasets were generated.

256 In these randomised datasets the correlation between the sediment yield and catchment
257 area can be calculated and the distribution of gradients and intercepts compared to that for the
258 correlation in the original dataset. If there is no significant difference between the best-fit line
259 for the observed and that from the randomised data, then any line fitted to the observed data
260 can be dismissed as spurious due to self-correlation. This approach differs from that of
261 Vickers et al. (2009) because in this case no distribution of the data is assumed while Vickers
262 et al. (2009) assumed a normal distribution and let that govern the randomisation process. In
263 this study it is only linear relationships in log-log space that were considered and tested.

264 Given the potential issues of self-correlation in the traditional approaches to
265 understanding sediment delivery ratio this study considers two alternative approaches. Firstly,
266 an alternative approach to understanding the change in sediment yield with catchment area is
267 not to consider the change in yield with catchment area but instead to consider the
268 incremental change in yield with catchment area, that is the sediment yield of the i^{th} km² in
269 the catchment that is required to meet the change in sediment flux observed as catchment area
270 increases. Secondly, the change in suspended sediment flux with catchment rather than the

271 change in sediment yield with catchment area was considered. The function of changing
272 suspended sediment flux with catchment area was assessed using a range of sigmoidal
273 functions (including logistic, hyperbolic, Weibull and Gompertz functions).

274

275

276 **3 Results**

277 For catchment-year combinations that met the criteria of sampling frequency ($f \leq 1$ month,
278 the bias-corrected suspended sediment yields have a median of 22.2 tonnes/km²/yr with a 5th
279 percentile of 5.4 tonnes/km²/yr and 95th percentile of 107.7 tonnes/km²/yr. Comparing the
280 sediment yield to the catchment area shows the expected decline and, perhaps not
281 surprisingly, the sediment yield declines with increasing catchment area (Figure 2).
282 Replotting Figure 2 without log-transformation shows that sediment yield declines
283 dramatically across the range 0 to 100 km² compared to the change in catchment size from
284 100 to 10000 km² (Figure 3). This implies that most sediment deposition happens in the
285 headwaters and not in the lower courses of a river.

286

287 *3.1 Sediment trends*

288 Figure 2 shows that there is not one trend of declining yield with area, but rather two
289 trends that bound the study catchments. The Trends (A and B – Figure 2) were derived from a
290 visual selection of the lowest bounding data points across all catchment areas in the dataset
291 and the highest bounding values in the dataset, and then plotting the best fit line through
292 them: note that point O was included in both the highest and lowest bounding datasets.
293 Trends A and B are distinguished by their difference in the rate of decline of yield with
294 increasing catchment area but both have a common point at point O (area 32 km², yield 374 ±
295 130 tonnes /km²/yr). Note that if Trends A and B were calculated without inclusion of point

296 O then the intersection of Trend A and B would be predicted at area = 29 km² and yield =
297 441.6 tonnes/km²/yr. Trend B was taken as only applying for catchment areas > 30 km² while
298 Trend A applies across the entire range. The equations of the bounding trends are:

299

300 Trend A

301 $\log_{10}(SS_{yield}) = 3.29 - 0.44\log_{10}(Area)$ (i)

302

303 Trend B

304 $\log_{10}(SS_{yield}) = 6.5 - 2.63\log_{10}(Area)$ (ii)

305

306 where: SS_{yield} = suspended sediment yield (tonnes/km²/yr); and Area = the catchment area
307 (km²).

308 The projection of Trend B beyond point O would give incredibly large values of erosion for
309 small catchments, but Trend A can be extrapolated to the y-axis. Trend A implies a
310 suspended sediment yield of 1950 tonnes/km²/yr at source, where source is taken as the y-
311 intercept value of the trend at 1 km². It is, therefore, possible to estimate an SDR for Trend A
312 (Figure 2). Given Equation (i), the SDR for the largest catchment (River Thames – 9948 km²)
313 is estimated to be 1.8%, but for some catchments represented by Equation (ii) the SDR would
314 be as low as 0.1%. At a 1.8% delivery ratio and a source suspended sediment export of 1950
315 tonnes/km²/yr, then 466 Mtonnes/yr of sediment has be retained in the catchment or turned
316 over to the atmosphere – we would suggest that this is improbably large amount of sediment
317 to be stored, processed or disposed each year in these UK catchments.

318 The value of 1950 tonnes/km²/yr as source term would appear large in comparison to
319 published values for the UK. Walling et al. (2002) working in 1.5 km² and 3.6 km²
320 agricultural catchments on mineral soils in England showed soil erosion rates up to 466

321 tonnes/km²/yr with 80% removal, or deposition, by the outlet of the catchment – a SDR of
322 20%. Defra (2005b) gave median values of net soil loss from arable fields in England as 410
323 tonnes/km²/yr and from English grasslands as 60 tonnes/km²/yr. Worrall et al. (2011) gave a
324 value of 406 tonnes/km²/yr for a bare peat plot in the South Pennines. Worrall et al. (2013)
325 suggested source suspended sediment values between 241 and 825 tonnes/km²/yr depending
326 upon the nature of the land use and soil types within the catchment with higher values for
327 organic soils under grass. Even though the value of 1950 tonnes/km²/yr would be the lower
328 value predicted from Equations (i) and (ii). There is little evidence, therefore, to support such
329 a high source value.

330 The unreasonably high value of the suspended sediment flux at source predicted by
331 Trend A compared to field observations from the literature does suggest that at small
332 catchment sizes the SDR is not linear with catchment area in log-log space and is only linear
333 over a mid-range of catchment sizes. Nevertheless, the data from the smallest catchment
334 available from the HMS dataset was for a 4 km² subcatchment of the River Dearne and its
335 yield is consistent with Trend A, i.e. even at a scale of 4 km² the sediment yield is 1229
336 tonnes/km²/yr (12.29 tonnes/ha/yr). Non-linear relationships in log-log space for sediment
337 yield and catchment area have been noted by, for example, de Vente and Poessen (2005).

338 Because trends identified in Equations (i) and (ii) are bounding trends, it is then
339 possible to express all results as a mixture of the bounding trends. Using the available
340 catchment characteristics, and stepwise regression, it was then possible to explore what
341 controlled a catchment's position between the two bounding trends for which the best-fit line
342 for the proportion of Trend A for a given catchment (ϕ_{trendA}) was:

343

$$344 \quad \phi_{\text{trendA}} = 0.51 + 0.29 \log_{10} \text{Area} - 0.12 \log_{10} \text{Arable} + 0.09 \log_{10} \text{Grass} + 0.08 \log_{10} \text{Orgmin} - 0.27 \log_{10} \text{AAR} \quad (\text{iii})$$

$$345 \quad r^2 = 0.49, n = 192$$

346

347 Where: Arable = the area of arable land use in the catchment (km²); Grass = the area of grass
348 land use in the catchment (km²); Orgmin = the area of organo-mineral soils in the catchment
349 (km²); and SAAR = the standard average annual rainfall (mm). Only terms found to be
350 significantly different from zero, at P < 0.05, were included in Equation (iii). Equation (iii) is
351 not proposed as a predictive equation, rather that there is a physical explanation for the
352 trends shown in Figure 2. Equation (iii) implies that larger catchments dominated by grass
353 and organo-mineral soils are more like Trend A while those catchments with a greater arable
354 land use and higher standard annual average rainfall are more akin to Trend B.

355 If the bounding trends to the sediment yield vs. catchment area curve are defined by
356 change with catchment area (Equations (i) and (ii)) but the position between two bounding
357 trends is controlled by nature of each catchment (Equation (iii)), then the position within the
358 defined space is defined by an interaction between catchment area and its land use and soil
359 types with a form of equation as:

360

$$361 \log_{10}SS_{yield} = k_1 + k_2f(\text{land use, soil type, hydroclimate})Area + k_3\log_{10}Area \quad (\text{iv})$$

362

363 Fitting this function to the available data gives a fit that can represent the bounding trends and
364 extreme values although the RMSE is 0.46 which suggests the fit for individual catchments
365 would be poor (Figure 4). Equation (iv) shows that land use, soils or hydrology becomes
366 more or less important as catchment area increases or decreases. Although many studies have
367 noted a variation in the relationship between sediment yield and catchment, very few have
368 explained this variation as shown above. Milliman and Syvitski (1992) did use maximum
369 altitude as well as catchment area to improve relationships but described land use, climate

370 and geology as of secondary importance – a result not supported here, although the study of
371 Milliman and Syvitshi (1992) was for larger catchments than any considered in this study.

372 The comparison of the nature of Trend A and Trend B suggests that, while catchments
373 following along Trend A have increasing suspended sediment flux with increasing catchment
374 area, for those catchments plotting along Trend B, the suspended sediment flux actually
375 decreases with increasing catchment area beyond point O (Figure 2). There are several
376 implications of this observation. Firstly, those catchments following Trend B are net
377 depositing catchments where the maximum sediment flux is that defined by point O (11955
378 tonnes/yr) and that this value would decrease for larger catchments. This implies that there
379 would be a catchment where almost no sediment was actually exported - in this case it would
380 be a catchment with area 616 km² (it should be noted that such a catchment along this trend
381 does not exist in this dataset). Therefore, just as for the small catchments where a linear
382 response gave unreasonable high values of yield, then the implication of a catchment which
383 has zero yield may also be unreasonable and an asymptote to the x-axis may be physically
384 preferable.

385 Secondly, an implication of the rate of decline of yield bounded by the two identified
386 trends, one of which implies increasing suspended sediment flux and the other implying a
387 decrease in suspended sediment flux with area catchment, suggests there is a boundary
388 between these two trends for which, although the yield declines, the actual flux does not. That
389 is, there is a boundary discernible below which suspended sediment flux decreases with
390 increasing catchment area, and above which it increases with increasing catchment area.
391 Given the point O is common to both trends then it is easy to define the line from which the
392 flux remains constant with increasing catchment area, ie. where the sediment flux remains the
393 same as that estimated for the catchment representing point O (River Dee at Mary Culter
394 Bridge 11955 tonnes/yr) across the range of catchment areas in the study dataset (Trend C -

395 Figure 2). All catchments can be defined relative to this line as being a net decreasing or a net
396 increasing catchment. Logistic regression can be used to assess what controls the membership
397 of these two groups. Here the best-fit logistic regression was:

398

$$399 \ln\left(\frac{\theta}{1-\theta}\right) = 8.8 - 0.012Evap_{act} - 0.0035Org - 0.006Grass \quad (v)$$

400 n=192

401

402 Where: θ = the probability of being a net increasing catchment; $Evap_{act}$ = the average actual
403 evaporation of the catchment (mm/yr); and Org = area of organic soils in the catchment
404 (km^2). The equation was 85% concordant with the data. The odds ratio suggests that none of
405 the variables included are more important than any other. The equation implies that a
406 catchment is more likely to be a net decreasing catchment if it is dominated by grassland and
407 organic soils and high actual evaporation. It should be noted that high actual evaporation
408 often correlates with higher rainfall rather than being indicative of drier catchments.

409 Since Trend C identifies no change in flux with increasing catchment area, this
410 implies that such catchments are experiencing a form of equilibrium in that, relative to point
411 O, they deliver to the outlet the same amount of sediment as enters the channel system.
412 Therefore, it would be reasonable to describe this situation as having a delivery ratio of 1
413 because, even if the sediment yield decreased, all the flux that entered the catchment was
414 delivered to the outlet. Often studies have used a decline in sediment yield with catchment
415 area to describe catchments as inefficient (Hinderer, 2012, Fryirs, 2013) whereas it could be
416 reasonable to describe catchments on these trends as very efficient as they export what comes
417 in to them.

418 We use our words carefully as it is important to stress that it is often the interpretation
419 of the SDR that we believe has been unhelpful and not the comparison of exports through a

420 catchment. When we describe catchments close to Trend C as efficient because they export
421 all the sediment that enters them, this is not to say that they convey the same sediment right
422 along their channel system but rather the same amount of sediment. These channels could
423 therefore be in a steady-state equilibrium that makes them appear as a perfect conduit of
424 sediment. Note that the back projection of Trend C to catchments sizes smaller than that
425 represented by point O would again suggest unrealistically high sediment yields and so again
426 there must be a curving to lower values so that the assumption of linearity is only true within
427 a constrained range of catchment sizes.

428

429 3.2 *Effects of self-correlation*

430 The test for self-correlation shows that the distribution of self-correlated regression
431 lines covers a considerable proportion of the results of the sediment yield against catchment
432 area (Figure 5). However, both Trend A and Trend B both lie outside the significant range of
433 the self-correlated random datasets and would therefore be considered as significantly
434 different from the self-correlated result.

435 The identification of self-correlation leads to some important results. Firstly, there is a
436 reasonable probability that any regression line on a graph of sediment yield against catchment
437 area will be spurious, i.e. not significantly different from results that would be expected from
438 the self-correlated nature of the comparison. Secondly, the self-correlation test does confirm
439 that there are bounding trends to the sediment yield for UK catchments and given that one of
440 these represents net deposition with increasing catchment area (Trend B) and the other
441 represents increasing sediment flux with increasing catchment area (Trend A), there must be
442 a point between these two trends where there is no change with catchment area and so the
443 result illustrated by Trend C does have a physical basis.

444 Considering not the yield with catchment area but the change in yield required from
445 the i^{th} km^2 of the catchment for the catchments along Trends A and B. The graph of the yield
446 of the i^{th} km^2 with increasing catchment area shows that for Trend A the yield declines
447 following a power law while that for Trend B shows a far more complex pattern that cannot
448 readily be described by a power law (Figure 6). A power law description can be equated with
449 a hypsometric curve for a catchment, i.e. slope change through a catchment is commonly
450 described by a power law and it is reasonable to suggest that sediment yield is a function of
451 slope among other things. If the change in sediment yield can be described by a power law
452 akin to a slope descriptor, we may propose that sediment yield change through a catchment is
453 not a matter of delivery but rather a matter of changing sediment supply through the
454 catchment, i.e. if slope through a catchment declines then each new km^2 of the catchment has
455 a lower capacity to produce sediment into the stream network.

456 For the River Thames catchment (9948 km^2) Trend A would now be interpreted as
457 having a sediment yield of $1950 \text{ tonnes/km}^2/\text{yr}$ in the first 1km^2 of the catchment and a
458 sediment yield of $18.5 \text{ tonnes /km}^2/\text{yr}$ in the very last km^2 of the catchment before the
459 monitoring point while the average sediment yield for the Thames catchment to 9948 km^2
460 point was $36.5 \text{ tonnes/km}^2/\text{yr}$. Alternatively, for Trend B what is observed is a prediction that
461 the sediment yield from the i^{th} km^2 would rapidly become negative and there would have to
462 be net deposition with this first occurring for the 47^{th} km^2 of the catchment and peaking in the
463 65^{th} km^2 of the catchment after which the rate of deposition declines becoming asymptotic to
464 zero with further increases in catchment area. If the change of sediment yield for Trend A
465 followed a curve which was readily interpretable as declining sediment source as the
466 catchment grew larger, then a change which has net deposition is a catchment which is
467 delivering less and less sediment into the next part of the catchment. The implications are
468 reduced inputs from lower-angle hillslopes further down the catchment and reduced capacity

469 for conveyance in lower-gradient channels notwithstanding the higher stream velocities and
470 more efficient channel cross-sections. It may indeed be, for the Thames, that Schumm's
471 "transport zone" encompasses most of the channel network and that there is now very little
472 additional net storage of sediment in the main floodplain system.

473 Given the self-correlated nature of the relationship between catchment area and
474 sediment yield, it is reasonable to consider alternative approaches. A graph of suspended
475 sediment flux (as opposed to sediment yield) against catchment area is be compromised by
476 self-correlation. Figure 7 shows that, as expected, the sediment flux increases with increasing
477 catchment size in the majority of cases and that a single linear relationship, although
478 significant, does not capture the nature of the data; as before, the data may be better described
479 as bounded by two trends (Figure 7). Note that in this case the two trends bounding these data
480 do not quite correspond to those identified above (Figure 2 – Equations (i) and (ii)). The point
481 O from Figure 2 is given on Figure 7 and the sites that constitute Trends A and B show that
482 the upper bound in Figure 2 is that identified as Trend A above but Trend B curves down and
483 away from Trend A but does not constitute the lower trend observed in Figure 2, i.e. trends of
484 declining suspended sediment flux are possible within the bounds of the data projected on
485 Figure 7.

486

487 3.3 *Suspended sediment – a new interpretation*

488 If Figure 7 is replotted on a double log plot, it is observed that the plot of suspended
489 sediment flux against catchment area could be bounded by two straight lines. A straight line
490 response on a log-log plot implies that a sigmoidal function would describe the data.
491 Therefore, we propose the use of the Gompertz function which is one of a family of
492 sigmoidal functions, but unlike others based upon hyperbolic functions, the parameters of a
493 Gompertz function are physically interpretable. However, sigmoidal functions have a low

494 response for small values of x, in this case small catchment areas: this is not observed for UK
495 data and so a short range effect at low values of catchment area was included:

496

$$497 \quad SS_{flux} = \alpha e^{\beta e^{\gamma Area}} - \delta e^{-\varepsilon Area} \quad (vi)$$

498

499 Where: α , β , γ , δ and ε = constants with α = maximum asymptote (tonnes/yr); β = the y
500 displacement with respect to catchment area (km^2); γ = the growth rate (tonnes/ km^2 /yr); δ =
501 The maximum small area correction (tonnes/yr); and ε = the decay constant for the small area
502 correction ($1/\text{km}^2$). Equation (v) was fitted to the data for the bounding trends (as for Figure 2
503 the highest and lowest bound datapoints were selected and then the best-fit of Equation (vi)
504 estimated), for Figure 7 and the best-fit equations are:

505

506 Trend D

$$507 \quad SS_{flux} = 350000e^{-3.16e^{-0.00089Area}} - 14219e^{-0.061Area} \quad (vii)$$

508

509 Trend E

$$510 \quad SS_{flux} = 87652e^{-5.53e^{0.00035Area}} - 109e^{-1.132Area} \quad (viii)$$

511

512 The fit of these equations shows that a maximum asymptote is apparent in the data that
513 represent a maximum sediment flux, in one case 350 ktonnes/yr and the other 87.7
514 ktonnes/yr.

515 The approach taken here can be considered for individual catchments. Within the
516 dataset considered here, it was possible to identify two catchments where multiple
517 measurements across scales were available – the rivers Trent and Ouse (Figure 8). Results for
518 individual catchments were included into the analysis illustrated in Figure 2 because plotting

519 in log space can act to cramp data points together. When plotted as sediment yield vs.
520 catchment area then it can be observed that each catchment's sediment yield evolves along a
521 line away from Trend B towards Trend A but at larger catchment areas the trend in sediment
522 yield for each specific catchment changes to trend parallel to Trend A. So, in individual
523 catchments in this UK dataset we do see catchments that show negative and positive
524 relationships between sediment yield and catchment area depending upon catchment location.
525 This has been observed for a number of catchments (e.g. Yellow River – Jiongxin and
526 Yunxia, 2005). When plotted as suspended sediment flux against catchment area, it is
527 apparent that both catchments could be described by an equation of the form of Equation (v)
528 where the best-fit equations were;

529

530 Trent

$$531 \quad SS_{flux} = 57980e^{-4.896e^{-0.00088Area}} - 188e^{-0.102Area} \quad (ix)$$

532

533 Ouse

$$534 \quad SS_{flux} = 62000e^{-5.0e^{0.0016Area}} - 188e^{-0.102Area} \quad (x)$$

535

536 These observations suggest that catchments will evolve and trend in complex ways across the
537 space bounded by Trends A and B.

538

539 **4. Discussion**

540 Does the approach here solve the problems that have been discussed for the SDR
541 method of analysing suspended sediment flux? Partly. Describing the variation in sediment
542 yield across a catchment using a source-decay approach does describe some of the data and in
543 particular data close to Trend A. However, it would predict no net storage in the catchment
544 which is not possible but clearly it is also absurd to have no change in source with catchment

545 area. Furthermore, the data along Trend B clearly show that net deposition would have to
546 occur in some catchments. Therefore, the reality is some mix of varying supply and delivery
547 and future research could be used to understand and compare the distribution of the
548 magnitude of sediment sources with increasing catchment size compared to the observed
549 change in sediment yield through the catchment.

550 Studies in Australia and Canada (e.g. Church et al., 1999) have shown positive
551 relationships between sediment yield and catchment area, but for the UK sediment yield is
552 always observed to decrease with increasing catchment area. Negative correlations between
553 sediment yield and catchment area have been interpreted in terms of changing land use,
554 hydrology and climatic factors, just as was possible here for the UK but the magnitude of the
555 gradients does suggest that there are catchments in the UK where sediment flux actually
556 declines through the catchment. It should be noted that the catchments that lie along Trend B
557 only show this behaviour above a certain catchment size; for a catchment area of less than 30
558 km² the sediment flux would increase with increasing catchment area. After all, suspended
559 sediment flux cannot decline from zero but must decline from some initial value. In this case
560 there appears to be a point at approximately 30 km² at which point net deposition begins.
561 Why would this occur? It is easy to envisage catchments with distinct breaks in slope
562 distribution with high slopes at small scales and very low gradients at larger scales, e.g.
563 transition from mountains to a “piedmont” zone (cf Lawler et al., 1999). Such changes in the
564 sediment cascade with increasing catchment area are worthy of further research beyond the
565 scope of this study.

566 Equally, the difference between Trend A and Trend B implies that there are
567 catchments that have no net change in sediment flux across the catchment. This only occurs
568 once a certain catchment area and flux have been achieved and therefore this does not mean
569 that these catchments are at an equilibrium, rather that net change is not occurring in their

570 lower reaches. However, it should be noted that, whenever tested with the observations of
571 sediment yield across two actual catchments, both showed complex changes with increasing
572 catchment area not necessarily following any single trend. Either, catchments can switch
573 between regimes with increasing scale, or the plot of sediment yield against catchment area is
574 misleading given the uncertainty in the data and the self-correlated nature of the plot. This
575 therefore may be the big question – how, where and why do catchments shift from transport
576 to storage and over what length and time scales?

577 However, there should be a prior question of whether we would start from here? The
578 SDR approach is used because it matches the information available from soil erosion models,
579 originally the USLE approach (Wischmeier and Smith, 1965). Given that a plot of sediment
580 yield vs catchment area, however drawn, is difficult to meaningfully interpret and suffers
581 from self-correlation, then the focus of attention and interpretation should be the actual graph
582 of suspended sediment flux vs. catchment area. This study has proposed that the curve of the
583 suspended sediment flux can be best described by a sigmoidal function, specifically, a
584 modified Gompertz function. The graph of the suspended sediment flux shows that, as
585 catchments get larger, the growth in suspended flux declines and reaches a maximum
586 described by the asymptote. It should be noted that no observed catchment had a sediment
587 flux that actually equalled this maximum value, perhaps because UK catchments have
588 relatively small areas compared to large river basins globally. The y-displacement in the
589 Gompertz equation represents the position of the step and, taking this as the position of
590 maximum gradient, then for equation (vii) the step position is at 1500 km^2 while for equation
591 (viii) it is 5100 km^2 ; for the River Trent it is 1300 km^2 and for the Ouse the 1100 km^2 . As the
592 y-displacement is here interpreted as the inflexion point in the step function, it represents the
593 catchment area at which the rate of suspended sediment flux is no longer increasing. This
594 could readily have a physical interpretation but, relative to the sediment delivery ratio

595 approach this shows the greatest rate of change is not at small catchment areas but well into
596 the catchment and so in this sense studying sediment yield from hillslope source areas alone
597 is misleading.

598 The growth rate in equation (vi) or the maximum rate of change observed for equation
599 (vii) is 114 tonnes/km² while for equation (viii) it is 11.14 tonnes/km². For the Trent the
600 maximum rate of change is 18.8 tonnes/km² and for the Ouse the maximum rate of change is
601 36.4 tonnes/km². However, given the small-area correction in equation (vi), then the
602 maximum rate change at the very smallest catchment areas, of the order of 1,280 tonnes/km²
603 for equation (vii) and 110 tonnes/km² for equation (viii). The gradient of the equation (v)
604 represents the incremental sediment yield.

605 The approach cannot say anything about the turnover of the organic proportion of the
606 suspended sediment. Several studies have considered the fate of organic carbon through
607 rivers, and therefore the impact of rivers on concentration of atmospheric greenhouses gases
608 (Battin et al., 2009). These studies have progressively developed methods for understanding
609 each component of this problem for the UK to understand the fate of fluvial organic matter
610 and the release of greenhouses gases including CO₂, CH₄ and N₂O (Worrall et al., 2007,
611 2012a,b) . Future research will focus upon reproducing this study for the organic matter flux
612 and then compare these results to those already produced for suspended sediment and so
613 assess the change in proportion of organic matter with catchment area relative to changes in
614 suspended sediment flux. We must always be mindful that change in POM flux and its
615 proportion across a catchment may be due to fractionation (i.e. not related to turnover) caused
616 by differential deposition, flocculation or transport of suspended sediment dependent upon its
617 organic matter content. Indeed, as source strength wanes across a catchment, slope, channel
618 or both, so does the proportion of organic matter, for example in the UK it would be common
619 to have peat soils in a headwater but not at the tidal limit of a catchment.

620

621 **5. Conclusion**

622 This study shows that the use of the sediment delivery ratio causes a number of mis-
623 interpretations. The SDR approach can be shown to suffer from a number of problems:

624 i) The graphical representation of sediment yield contains trends that encapsulate
625 both increasing, decreasing and constant sediment flux with increasing catchment
626 area. Some catchments showed perfect “delivery” of sediment whilst in others
627 sediment yield declined with increasing catchment area.

628 ii) The relationship of SDR against catchment area cannot be linear over the range of
629 catchment sizes.

630 iii) The relationship of SDR against catchment area suffers from considerable self-
631 correlation and any relationship derived from the graph should be tested before use or
632 interpretation.

633 iv) Alternatively, a plot of suspended sediment flux against catchment area does not
634 suffer from self-correlation and can be readily described by a simple empirical
635 function based upon the combination of a Gompertz function and a short-range,
636 small-area correction based upon an exponential decay function.

637 This study can find no reason why the SDR approach should continue to be employed in
638 the form that it has been traditionally been used.

639

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643

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761 Figure 1. The location of the catchments used in this study for which suspended sediment flux could
762 be calculated.

763

764 Figure 2. The log-log plot of the average suspended sediment yield against catchment area for all
765 study catchments.

766

767 Figure 3. The average suspended sediment yield against catchment area for all study catchments
768 showing that the suspended sediment yield changes most rapidly for the smallest catchments.

769

770 Figure 4. Fit of Equation (iv) to the suspended sediment yield against catchment area.

771

772 Figure 5. The best-fit self-correlated line and its 95% confidence interval in comparison to the annual
773 average suspended sediment yield against catchment area.

774

775 Figure 6. The changes in incremental suspended sediment yield with catchment area for a) Trend A;
776 and b) Trend B (Figure 2).

777

778 Figure 7. The average suspended sediment flux with catchment area for all study catchments.

779

780 Figure 8. The change in suspended sediment for sites through the catchments of the Rivers Trent and
781 Ouse against their catchment area.

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